Article 77(3)(c) Request Supporting Information Report

EVALUATION OF ENDOCRINE DISRUPTING PROPERTIES FOR THE ENVIRONMENT

Substance Name: 4,4'-methylenediphenol ("Bisphenol F")

EC Number: 210-658-2

CAS Number: 620-92-8

Submitted by: Germany

Date: April 2022

CONTENTS

	_
ABBREVIATIONS	
BACKGROUND	4
RESULTS OF EVALUATION OF ENDOCRINE DISRUPTING PROPERTIES FOR THE ENVIRONMENT	
1. IDENTITY OF THE SUBSTANCE AND PHYSICAL AND CHEMICAL PROPERTIES	7
1.1. Name and other identifiers of the substance	. 7
2. HARMONISED CLASSIFICATION AND LABELLING	9
3. ENVIRONMENTAL HAZARD ASSESSMENT - ENDOCRINE DISRUPTION	10
3.1. General approach	10 11 14 20 21 21 21 22 22
TABLES Table 1: Substance identity Table 2: Structurally related substance(s) identity Table 3: Constituents of structurally related substance EC 908-912-9 Table 4: Overview of physicochemical properties Table 5: Available in vitro data on BPF Table 6: Available in vivo data with fish on BPF	. 8 . 9 . 9 . 1

ABBREVIATIONS

AhR Aryl-hydrocarbon receptor

AR Androgen receptor

ARN Assessment of Regulatory Needs

BPA Bisphenol A (EC 201-245-8, CAS 80-05-7)
BPAF Bisphenol AF (EC 216-036-7, CAS 1478-61-1)
BPB Bisphenol B (EC 201-025-1, CAS 77-40-7)
BPF Bisphenol F (EC 210-658-2, CAS 620-92-8)
BPS Bisphenol S (EC 201-250-5, CAS 80-09-1)

DHT Eihydrotestosterone dpf days post-fertilisation

E2 Estradiol

EAG Expert Advisory Group
EC Effective concentration
ECHA European Chemicals Agency

ED Endocrine Disruptor

EDC Endocrine Disrupting Compound EFSA European Food Safety Authority

ER Estrogen receptor
hpf hours post-fertilisation
GSI Gonadosomatic index
HSI Hepatosomatic index

HPT Hypothalamic-pituitary-thyroid

IC Inhibitory concentration JRC Joint Research Center LC Lethal concentration

LC-MS Liquid chromatography/mass spectroscopy

LDH Lactate dehydrogenase

LOEC Lowest Observed Effect Concentration

MoA Mode of action

MSC Member State Committee

(MS)CA (Member State) Competent Authority

NF Nieuwkoop and Faber (developmental staging for xenopus laevis)

NO(A)EL no observed (adverse) effect level

OECD Organisation for Economic Co-operation and Development

(q)PCR (quantitative) Polymerase chain reaction

RAC Committee for Risk Assessment RMOA Risk Management Option Analysis

SEv Substance Evaluation

SVHC Substance of very high concern

T3 Triiodothyronine

T4 Thyroxine

TH Thyroid hormone

TR Thyroid hormone receptor

TTR Transthyretin VTG Vitellogenin

WHO World Health Organisation

WoE Weight of Evidence

Background

The German Competent Authority (DE CA) has indicated through an entry in the Registry of Intentions that they plan to submit a proposal to restrict 4,4'-isopropylidenediphenol (EC 201-245-8, CAS 80-05-7, Bisphenol A) and bisphenols exhibiting a similar concern for the environment from uses which lead to emissions to the environment.

Following submission of the restriction dossier and a positive conformity check, the hazards and risks of the proposal would be evaluated by ECHA's Risk Assessment Committee (RAC).

In addition to bisphenols with confirmed endocrine disrupting properties for the environment, the restriction proposal is planned to also include the following two bisphenols for which no EU-wide consensus has so far been reached on their endocrine disrupting properties for the environment:

- 4,4'-methylenediphenol (EC 210-658-2, CAS 620-92-8, "Bisphenol F")
- 4,4'-[2,2,2-trifluoro-1-(trifluoromethyl)ethylidene]diphenol (EC 216-036-7, CAS 1478-61-1, "Bisphenol AF") and its salts

Generally, EU-wide opinion forming on whether a substance has endocrine disrupting properties for the environment is achieved through its identification as a Substance of Very High Concern (SVHC), which involves opinion making by the ECHA Member State Committee (MSC). MSC has extensive previous experience of such evaluations. To avoid a situation where RAC is required to conclude on whether a substance fulfils the criteria for endocrine disruption for the environment, rather than the MSC, it is anticipated that ECHA will facilitate an ad-hoc process under Article 77(3)(c) of REACH to enable the MSC to give an opinion on the endocrine disrupting properties of the substances identified above, which can then be used by RAC as the basis of their evaluation. This assessment report is submitted to ECHA by the DE CA to be used as the basis for the opinion-making of the MSC.

RESULTS OF EVALUATION OF ENDOCRINE DISRUPTING PROPERTIES FOR THE ENVIRONMENT

Substance name(s): 4,4'-methylenediphenol ("BPF")

EC number(s): 210-658-2

CAS number(s): 620-92-8

The substance acts as an endocrine disruptor (ED) in the environment according to the ED criteria of the World Health Organisation (WHO).

Summary of how the substance meets the criteria for an environmental ED:

4,4'-methylenediphenol (Bisphenol F, BPF) acts as an ED in the environment based on available information from *in vitro* and *in vivo* data in fish:

Adverse effects relevant for ED identification

The available data for BPF show clear and consistent adverse and population relevant effects on reproduction and sexual development in two zebrafish studies. The observed effects fit to an estrogenic and/or anti-androgenic mode of action and no indications of further non-ED mediated pathways were found in the two studies.

Endocrine activity

The available *in vitro* as well as *in vivo* mechanistic data clearly and consistently demonstrate an estrogenic and anti-androgenic activity of BPF.

Additionally, two *in vitro* studies also show an interference with testosterone production in cellular assays.

Plausible link between adverse effects and endocrine activity

BPF may have multiple modes of endocrine action (estrogenic, anti-androgenic, thyroidal activity and interference with steroidogenesis) that might interact and are difficult to distinguish from each other. However, the estrogenic and/or anti-androgenic effects of BPF in fish are consistently observed in the available studies showing significant effects on egg production, hatching and survival of F1 larvae, sex ratio and gonadal development. Estrogenic and anti-androgenic modes of action are well known to be involved in the regulation of sexual development and reproduction. Considering the observed concomitant decrease in plasma testosterone levels and the increase in plasma estradiol levels as well as the increase in vitellogenin (VTG) levels and gene expression in male fish, the link between these endocrine activities and the observed adverse effects on fish is highly plausible.

Other supporting evidence

In vitro and in vivo data from fish and amphibians point to an interference of BPF with the hypothalamic-pituitary-thyroid (HPT) axis.

Additionally, the available human health data support the conclusion for an estrogenic and/or antiandrogenic activity of BPF, even though the reliability of these studies have not

been assessed here. Studies on rats show consistently a decrease in serum testosterone levels and a decrease in sperm motility in the offspring of treated female rats. An increase in uterus weight was observed in juvenile rats.

The link between the observed effects and the specific estrogenic and/or anti-androgenic activity of BPF is further supported by the analogy of BPF to BPA and BPA. The data available for BPA and BPB, both of which share very similar chemicals structures compared to BPF, show well defined adverse effects and modes of action that fit to an estrogenic mode of action in fish. Based on these data BPA and BPB have been identified already as SVHC due to its endocrine disrupting properties in the environment.

Conclusion on ED properties

Overall, BPF has estrogen agonistic properties and induces adverse effects in fish that are plausibly mediated by this endocrine activity.

Furthermore, *in vivo* and *in vitro* evidence shows that BPF has androgen antagonistic properties. This endocrine activity could also plausibly contribute to the observed adverse effects on reproduction and sexual development in fish.

The effects observed in fish are relevant for the environment as an effect on the reproductive function and the sexual development can have consequences at a population level.

Therefore, there is scientific evidence to conclude that BPF fulfils the definition of an endocrine disruptor in the environment.

Registration dossiers submitted for the substance: No

1. Identity of the substance and physical and chemical properties

1.1. Name and other identifiers of the substance

Table 1: Substance identity

EC number:	210-658-2
EC name:	4,4'-methylenediphenol
SMILES:	Oc2ccc(Cc1ccc(O)cc1)cc2
CAS number:	620-92-8
IUPAC name:	4,4'-Dihydroxydiphenylmethane
Index number in Annex VI of the CLP Regulation	Not available
Molecular formula:	$C_{13}H_{12}O_2$
Molecular weight range:	200.20 g/mol
Synonyms:	Bisphenol F BPF

Substance type: mono-constituent

Structural formula:

1.2. Identity and composition of structurally related substances

In lieu of disseminated data from a registration dossier on BPF itself, the following data is reported from the ECHA dissemination database on EC 908-921-9, the multi-constituent substance comprising BPF and its two stereoisomers, section "Physical & chemical properties".1

¹ Via https://echa.europa.eu/de/registration-dossier/-/registered-dossier/26309 accessed 25 April 2022

Table 2: Structurally related substance(s) identity

EC number:	908-912-9
EC name:	2-[(2-hydroxyphenyl)methyl]phenol; 2-[(4-hydroxyphenyl)methyl]phenol; 4-[(4-hydroxyphenyl)methyl]phenol
SMILES:	See below
CAS number (in the EC inventory):	1333-16-0
CAS number:	1333-16-0
IUPAC name:	Reaction mass of 2,2'-methylenediphenol and 4,4'-methylenediphenol and o-[(4-hydroxyphenyl)methyl]phenol
Index number in Annex VI of the CLI Regulation	N/A
Molecular formula:	$C_{13}H_{12}O_2$
Molecular weight range:	200.20 g/mol
Synonyms (trade names):	GX-460 KD-9005 KD-9007 KD-9009 KDF-214M KDF-438 KDN-253 SP-2000 SP-2000P YDPN-638A80 YDPN-638N

Substance type: multi-constituent

Structurally related substance(s) formula:

The registered substance is an isomeric mixture of (from left to right) 4,4'-BPF, 2,4'-BPF and 2,2'-BPF (see below).

Table 3: Constituents of structurally related substance EC 908-912-9

Constituents	EC Number	CAS number	SMILES / structural formula
2,2'-methylenediphenol	219-578-2	2467-02-9	Oc1ccccc1Cc2cccc2O OH OH
4,4'-methylenediphenol	210-658-2	620-92-8	Oc2ccc(Cc1ccc(O)cc1)cc2
o-[(4- hydroxyphenyl)methyl]phenol	219-579-8	2467-03-0	Oc2ccc(Cc1ccccc1O)cc2 OH HO

1.3. Physicochemical properties

Table 4: Overview of physicochemical properties

Property	Value [Unit]	Reference/source of information
Physical state at 20°C and 101.3 kPa	Crystalline, light red	Safety data sheet (Sigma- Aldrich 2022) ²
Melting/freezing point	162-164 °C	Safety data sheet (Sigma- Aldrich 2022) ²
Boiling point	291 °C @ 101.3 kPa	OECD 103 (Test substance: EC 908-912-9) ¹
Vapour pressure	0 Pa @ 20-25 °C	OECD 104 (Test substance: EC 908-912-9) ¹
Density	1.279 g/cm³ @ 20 °C	OECD 109 (Test substance: EC 908-912-9) ¹
Water solubility	2.97 g/L @ 20 °C @ pH 6.2- 6.3	OECD 105 (Test substance: EC 908-912-9) ¹
Partition coefficient n- octanol/water (log value)	log P _{ow} = 2.91 @ 20 °C	OECD 107 (Test substance: EC 908-912-9) ¹

2. Harmonised classification and labelling

BPF is not covered by an Index number in part 3 of Annex VI to the CLP Regulation.

² https://www.sigmaaldrich.com/DE/en/sds/ALDRICH/B47006

3. Environmental hazard assessment – endocrine disruption

3.1. General approach

The available data for BPF is presented with the focus on tests and endpoints that can be conclusive for their endocrine disrupting properties in the environment. Similarity of concern to BPA and BPB with respect to intrinsic hazardous properties in this case is considered to be met if based on the available data the substance fulfils all of the following criteria in accordance with the WHO/IPCS definition of an endocrine disruptor in the environment as interpreted by the EC ED EAG (JRC, 2013):

- Show an adverse and population relevant effect in organisms. With regard to this, all effects that impact the survival, the growth or the reproduction of an organism are considered to be adverse and of population relevance.
- Show an endocrine activity; and
- An endocrine mode of action (MoA), i.e. there is a biologically plausible link between the endocrine activity and the adverse effects observed.

To conclude if these criteria are fulfilled, a weight of evidence approach is used. All relevant and reliable data available when compiling this dossier are considered. This comprises *in silico*, in vitro and in vivo data from standard as well as from non-standard exploratory studies. The data were grouped into three categories following the conceptual framework of the OECD Revised Guidance Document 150 (OECD 2018) and the EU EDC guidance (ECHA & EFSA, 2018): I) *in vitro* and *ex vivo* mechanistic parameters, II) in vivo mechanistic parameters and III) parameters providing information on adversity. Based on the adverse effects identified, results were further integrated into lines of evidence, defined as a "set of relevant information grouped to assess a hypothesis," using a weight-of-evidence approach (ECHA & EFSA, 2018).

The relevance of a given study is assumed if the study design allows answering questions regarding an endocrine mode of action and/or regarding adverse effects that are or can be mediated via an endocrine MoA. To judge on the reliability of data, all studies were assessed according to the Klimisch scoring system (Klimisch et al., 1997).

3.2. Previous assessments of the endocrine properties of BPF

In March 2021 the SE CA provided an RMOA for BPF also discussing the ED concern with a focus on ED properties for human health. The SE CA concluded that the current evidence might not be strong enough to be used as basis for SVHC identification (ED HH) but might be used as the basis for a read-across to BPA to propose a classification of BPF as Repr. 1B. Evaluation of ED properties for the environment was not the main focus of the SE CA's RMOA.³

3.3. Assessment of information on ED properties of BPF

A literature search was performed to identify all studies available in open literature relevant for the assessment of endocrine disrupting properties of Bisphenol F with respect to the environment. As the substance itself is not registered under REACH, an assessment

³ Risk Management Option Analysis Conclusion Document on 4,4'-methylidenediphenol (Bisphenol F) dated 24 March 2021: https://echa.europa.eu/documents/10162/785a51c1-3391-1b81-e532-cc83b665ea60

of the results presented in study reports from the registration dossier could not be included in the assessment here.

The last PubMed search was performed on September 10th 2021. This resulted in 498 hits. The respective search strings and number of hits are presented in the table below.

Date of search	Database	Search string	Number of articles	Comment
Again on 10.09.2021	PubMed	(Bisphenol F) OR (4,4'- methylenediphenyl) OR (620-92-8 OR (210-658-2) AND ((fish OR human OR toxicity) OR (endocrine* OR hormone* OR androgen* OR estrogen* OR thyroid* OR steroid*))	498 publications Manual screening narrowed the result to 36 potentially relevant publications	The title and abstracts of the identified articles were screened manually. Expert judgement was also used to determine whether a study was likely to provide information of relevance for the ED assessment.

The quality and reliability of all in vitro and in vivo studies were assessed, and each study was assigned a reliability score based on the Klimisch categories 1, 2, 3 or 4 (Klimisch et al., 1997) combined with an expert judgement statement.

The following sections report and discuss the relevant and reliable data to conclude on the environmental ED properties of BPF according to the approach described above.

3.3.1. In vitro information indicative of endocrine activity

There are eleven studies available investigating estrogenic, anti-estrogenic, androgenic, anti-androgenic, thyroidal and steroidal activity of BPF. Additionally, growth hormone production, metabolism and cytotoxicity were investigated in some studies. All studies are rated as Klimisch 2, except the study performed by Park et al. (2020), which is rated Klimisch 1 as it follows OECD TG 455.

The following table provides an overview of the available in vitro data which are further discussed in the respective subsections.

Table 5: Available in vitro data on BPF

Method	Results	Reference
Culture system	BPA: 10 nM reduces basal testosterone secretion of	Eladak, et
(foetal testis assay)	human foetal testis explants	al., 2015
tested with BPS, BPF	Addition of LH in FeTA system enhances BPA min	
	effective conc in mouse and human but not in rat foetal testis	(Klimisch 2)
	BPS / BPF: 10 nM decreased basal testosterone secretion by human foetal testis	
	BPA and BPF: 1000 nM min effective concentration in foetal mouse testes	
	BPA / BPS / BPF: 10000 nM reduced <i>Insl3</i> expression in cultured mouse foetal testes	
Luciferase reporter	ERa-agonism: 63.1% max efficacy, rel. $EC_{50} = 1.6 \cdot 10^{-1}$	Pelch et al.,
gene assay with	⁶ M (BPA: 110.3%, 1.2·10 ⁻⁶ M)	2019
HepG2 cells	ERβ-agonism: 97.9 %max Efficacy, rel. $EC_{50} = 1.3 \cdot 10^{-1}$	
transiently	⁶ M (BPA: 97.4%, 3.5·10 ⁻⁷ M)	(Klimisch 2)
transfected with ERa or ERB	ERa-antagonism: 37.6% max. inhibition, rel. $EC_{50} = -$ (BPA: 11.9%, -)	
0. 5.4	ERβ-antagonism: 12.5 % max. inhibition, rel. EC ₅₀ = -	

_		1
	(BPA: 35.2%, -) AR antagonism: 66.2% max. inhibition, rel. $IC_{50} = 5.9 \cdot 10^{-6}M$ (BPA: 98.8%, $5.1 \cdot 10^{-7}M$)	
Mixture effects: Stably transfected transcriptional	ER transactivation activity in Hela9903 cells. HeLa9903 cells treated with E2 (1 nM) or: BPA: 1.0×10^{-10} to 1.0×10^{-4} M $\rightarrow 1.51$ µM compared to	Park et al., 2020
activation (STTA) assay according to OECD TG 455:	100% activity induced by E2 BPF: 1.0×10^{-10} to 1.0×10^{-4} M $\rightarrow 3.96 \mu$ M	(Klimisch 1)
Human ERa- expressing HeLa9903 cells + MCF-7 cells	ERa protein expression in MCF-7 cells MCF-7 cells were treated with E2 (10 nM) or: BPA: 1 μM decreased the Era-expression 0.69-fold BPF: 1 μM no change observed, 10 μM 0.53-fold reduced ERa protein levels (for BPA, BPF, BPS)	
	androgenic activities in AR-EcoScreen cells AR-Ecoscreen cells were treated with DHT (10 nM) and: BPA: $1.0 \cdot 10^{-10}$ to $1.0 \cdot 10^{-5}$ M; BPF: $1.0 \cdot 10^{-9}$ to $1.0 \cdot 10^{-5}$ M \rightarrow no AR agonist activity;	
	anti-androgenic activities in AR-EcoScreen cells AR-Ecoscreen cells were treated with DHT (500 pM) or/and HF (100 nM) and BPA $1.0\cdot10^{-10}$ to $1.0\cdot10^{-5}$ M, BPF $1.0\cdot10^{-9}$ to $1.0\cdot10^{-4}$ M BPA and BPF had antagonistic activity at $10~\mu\text{M}$ Compared with the 100% activity induced by 500 pM DHT (positive control), AR activity was decreased 50% by 8.93 μ M BPA, 10.8 μ M BPF.	
	AhR activities in DR-EcoScreen cells. DR-Ecoscreen cells were treated with TCDD (100pM) and BPA $1.0\cdot10^{-10}$ to $1.0\cdot10^{-5}$ M, BPF $1.0\cdot10^{-9}$ to $1.0\cdot10^{-4}$ M \rightarrow 10 μ M BPA, 100 μ M BPF, and 100 μ M BPS had very low, but significant, AhR-mediated activity	
Comparative study in vitro (and in vivo) sperms and testicular tissue of rats (BPA,	Antioxidant enzyme activities and oxidative stress markers were induced in the testes, whereas testosterone production was reduced	Ullah et al., 2018 (Klimisch 2)
BPB, BPF, BPS) doses (0, 1, 10, and 100 ng/ml) stock solutions with ethanol tissue from 7 Sprague–Dawley male adult rats		
In vitro (and in vivo) zebrafish	Efficient transactivation of all zebrafish estrogen receptor (zfER) subtypes in zebrafish hepatic reporter cell lines (ZELH-zFERs). BPA was selective for zfERa	Le Fol et al., 2017
	while BPS and BPF were slightly more potent on zfERβ subtypes	(Klimisch 2)
HPLC-APCI/APPI- HRMS was coupled with CALUX <i>in vitro</i>	TTR binding potency: BPAF = BPF > BPA = BPE anti-androgenic activities: BPAF > BPE > BPA > BPTMC > BPF > BPS	Šauer et al., 2021
reporter gene bioassays		(Klimisch 2)
Hepatocytes isolated from liver of O.mykiss	LDH assay: increasing cytotoxicity in cells (dosedependent)	Aykut and Kaptaner, 2021
24 h expo to BPF 0, 15.63, 31.25, 62.50, 125, 250, and 500 µM	Increased malondialdehyde content (indicative of lipid peroxidation) sign between 15.63 and 250 µM.	(Klimisch 2)

		1
LDH (lactate	Toxic mechanism mainly based on cell membrane	
dehydrogenase	damage and oxidative stress	
assay) and		
antioxidant defence		
system indicators		
2,2'-BPF (CAS 2467-	2) 4,4'-BPF slightly slower metabolization than BPA and	Punt et al.,
02-9); 4,4'-BPF (CAS	2,2'-BPF	2019
620-92-8)	2,2 511	2013
1) Caco-2 cells: in	Yeast estrogen bioassay	(Klimisch 2)
vitro intestinal	$EC_{50} (4,4'-BPF) = 20 \mu M$	(Killinisch Z)
transport study	EC_{50} (2,2'-BPF) = NA but anti-ER IC_{50} = 70 μ M	
2) HepaRG: in vitro	EC_{50} (BPA) = 20 μ M	
liver metabolism		
study	Yeast anti-androgen bioassay:	
3) QIVIVE for	$IC_{50} (4,4'-BPF) = 20 \mu M$	
EC ₅₀ /IC ₅₀	$IC_{50}(2,2'-BPF) = NA$	
extrapolation	IC_{50} (BPA) = 30 μ M	
fluorescence	Fluorescence competitive binding assay: BPS and BPF	Zhang et al.,
competitive binding	bound to TH receptors (TRa and TRβ) (binding	2018
assay	potencies an order of magnitude lower than BPA)	
Coactivator	Coactivator recruitment assay: BPS and BPF recruited	(Klimisch 2)
recruitment assay	coactivator to TRβ but not TRα, with weaker potencies	,
TR-mediated	than BPA	
luciferase reporter	TR-mediated reporter gene transcription assay:	
gene system using	agonistic actions in the absence or presence of T3	
GH3 cells	BPA, BPF, BPS induced TH-dependent GH3 cell	
T-screen assay	proliferation	
1-screen assay	BPA and BPF inhibited T3 induction in presence of T3	
In vitro		Kitamura et
	MCF-7 estrogen luciferase reporter assay:	
ERE-luciferase	BPF-EC ₅₀ = $1.0 \mu M$	al., 2005
reporter assay using	BPA-EC ₅₀ = $0.63 \mu M$	
MCF-7 cells + ARE-	BPS-EC ₅₀ = $1.1 \mu M$	(Klimisch 2)
luciferase reporter		
assay + induction of	Luciferase Reporter Assay (anti-androgenic act. against	
growth hormone	DHT):	
production in GH3	$BPF-IC_{50} = 12 \mu M$	
cells	BPA- $IC_{50} = 4.3 \mu M$	
	$BPS-IC_{50} = 17 \mu M$	
	Growth hormone production assay of GH3 cells:	
	BPA, BPF showed no activity	
<u> </u>	, , , , , , , , , , , , , , , , , , , ,	

3.1.1.1. Estrogen pathway

Several in vitro reporter gene and transactivation assays show a clear agonistic activity of BPF on both ER α and ER β subtypes of the receptor proteins (Kitamura et al., 2005, Le Fol et al., 2017, Pelch et al., 2019, Park et al., 2020). In these studies, BPF (i.e. 4,4 ´-BPF) showed similar EC $_{50}$ values in the low μ M range as BPA. Additionally, there is a yeast estrogen bioassay performed by Leuwen et al. (2019) which also shows estrogen agonistic activity of BPF and BPA, both at EC $_{50}$ values of 20 μ M.

Anti-estrogenic activities of BPF were observed in the studies by Pelch et al. (2019) and Park et al. (2020) in the transactivation assay set up and in an MCF-7 cell-based ERa protein expression level.

3.1.1.2. Androgen pathway

No androgen agonistic activity is reported in the available studies. Anti-androgenic activity was observed in several assays (Kitamura et al., 2005, Punt et al., 2019, Pelch et al., 2019, Park et al., 2020, Šauer et al., 2021). Reported IC₅₀ values and effect concentrations

of BPF and BPA were similar or at least in the same order of magnitude.

3.1.1.3. Thyroid pathway

BPF showed binding to both subtypes of the thyroid receptor protein in a competitive binding assay (Zhang et al., 2018). The same study found that BPS and BPF recruited coactivator to $TR\beta$ but not to $TR\alpha$ in a coactivator recruitment assay but with weaker potencies compared to BPA. Additionally, Zhang et al. (2018) performed a TR-mediated reporter gene transcription assay and found agonistic actions of BPF in the absence or presence of TR3. The study further showed that BPA, BPF, BPS induced TR4-dependent TR8-dependent TR9-dependent TR9-

3.1.1.4. Steroid pathway

Eladak et al. (2015) found that BPF and BPA comparably decreased basal testosterone secretion by human and mouse foetal testis in a culture assay set up. Growth hormone production was analysed in GH3 cells by Kitamura et al. (2005) without any effects of BPF and BPA.

Cytotoxicity of BPF in hepatocytes isolated from liver of *Oncorhynchus mykiss* was investigated by Aykut and Kaptaner (2021) using a LDH assay. This study observed significant cytotoxic effects based on oxidative stress (lipid peroxidation and membrane damage) between 15.63 and 250 μ M.

3.1.1.5. Conclusion

BPF shows significant endocrine activity in the available in vitro studies. Most prominent are estrogen agonistic effects and anti-androgenic effects with effect concentrations (in the low μ M range) in the same order of magnitude as or even equal to the concentrations observed for BPA, which is already identified as an ED in the environment based on estrogenic and or anti-androgenic effects in fish. These results are consistent across different cell lines and studies. BPF also shows thyroidal activity based on binding to the thyroid receptor protein as well as to the transport protein TTR. One study demonstrated an inhibitory effect of BPF on testosterone production in a testis cell culture assay. Significant cytotoxicity of BPF was observed in one study using isolated hepatocytes from rainbow trout, however severe effects were seen at higher concentrations compared to

effect concentrations triggering ED specific endpoints in the other available in vitro studies.

3.3.2. In vivo information indicative of endocrine activity

In the following table the available *in vivo* fish and amphibian data is summarised. All studies are rated as Klimisch 2 since they represent high quality exploratory studies.

Table 6: Available in vivo data with fish on BPF

Method	Results	Reference
Danio rerio (AB strain)	Geometric mean conc.:	Yang et al.,
28 ± 0.5 °C	0.00071, 0.0072, 0.079, 0.75 mg/L	2017
6 male and 6 female fish in 10-L glass	No mortalities in any treatment	
aquaria	HSI at 1 mg/L sign higher in males	(Klimisch 2)
control, 0.001, 0.01, 0.1, and 1 mg/L	+ GSI sign decreased in females	
BPF for 21 days following OECD	and males at 1 mg/L	
guidelines 229 and 230	Egg production, hatching rate,	
0.01% DMSO	survival rate sign reduced at 1 mg/L	
Two replicates	Exposing parental fish resulted in	

Semi-static Histological analysis under microscope Hormone measurements using ELISA RNA extraction and gene expression analysis	malformed embryos and larvae (sign at 0.1 and 1 mg/L) Histological examination: decrease in number of early sperm stages, enlargement of interstitial space in male fish, females: lower proportion of later follicular stages; no intersex individuals Sign differences in concentrations of steroid hormones: decrease in T conc. In homogenate of 0.001, 0.1 and 1 mg/L + sign increase of E2 conc. In 0.1 and 1 mg/L Liver: sign increased vtg1 expression in males (dosedependent), but not in females.	
Zebrafish (<i>Danio rerio</i> , AB strain) Acute tox test: 20 eggs (2 hpf) in 6- well cell culture plates with 10 mL test solution in triplicate (0, 1, 10, 100, 1000, 2000, 4000, 6000, or 8000 μg/L, 0.01% DMSO) expo: 2 hpf to 120 hpf 16 h light per day 28 ± 0.5 °C Long-term expo: 500 eggs in 500 mL glass beakers in triplicate (0, 1, 10, 100, or 1000 μg/L) Semi-static At 10 dpf larvae transferred to 10 L glass aquaria Expo: 1 to 60 dpf Gonadal histology (microscope) ELISA Rea-time PCR Analytical verification: LC-MS Statistics: Levene's test on homogeneity of variances + Kolmogorov-Smirnov test on normality; one-way analysis of variance Tukey's post -hoc test (p < 0.05)	Mean measured conc: 0.93, 11.51, 98.56, and 1047.93 μg/L 96h-LC ₅₀ = 10,030 μg/L 120h-LC ₅₀ = 9391 μg/L 100 and 1000 μg/L BPF: - led to trend for a female sex ratio bias - Gonad histology: 10 and 22 % intersex fish + all male fish developed ovo-testes - Abnormal testicular development Steroid hormone conc.: - Conc. T decreased sign in ≥ 10 μg/L - E2 increased sign in ≥ 10 μg/L VTG expression increased sign from 20 dpf in ≥ 10 μg/L expo groups	Yang et al., 2018 (Klimisch 2)
Zebrafish embryos (wild-type WIK strain) BPA, BPF, BPS, BPAF Ethanol 0.01%	96h-LC ₅₀ : BPA 12 mg/L BPF 32 mg/L BPS 199 mg/L	Moreman et al., 2017 (Klimisch 2)
LC-MS analysis Co-exposure with ICI 182780 (CAS 129453-61-8) (ER antagonist) 20 embryos per group in 100 mL water in triplicate 28 ± 1 °C Semi-static Toxic effects + morphological abnormalities: OECD TG 236 (96h) Estrogenic response by GFP induction 0 to 120 hpf Zebrafish (wild-type AB)	Hatching success 72hpf EC50: BPA 5.7 mg/L BPF 14 mg/L BPS 155 mg/L Similar morphological abnormalities, including acriac edema, spinal malformation and craniofacial abnormalities (96 hpf) + lack of pigmentation at 10 mg BPF/L and above All treatments with BPF:	Vuon et al
Control – solvent control - 0.0005 – 0.5 – 5.0 mg/L triplicates 2 hpf in 24-well plates (RNA-Seq Test)	Reduction of swimming distance and locomotive activity at 6 dpf; Activity range and swim speed of BPF-treated larvae generally lower	Yuan et al., 2019 (Klimisch 2)

OR QPCR 2 hpf in 1 L beaker (150 mL	than control	
exposure solution) with 120 embryos;		
at 48 hpe 85 embryos from each	0.5 and 5.0 mg BPF/L: sign affect	
replicate OR Chip-seq test	motor neuron development at 72	
	hpf with an inhibition of axon growth → comparative analysis: BPF	
	stronger influence on zebrafish	
	motor neuron development than	
	BPA	
	No transcriptional changes observed	
	for estrogenic genes in BPF	
Danio rerio (AB strain)	treatment OECD TG 212: $96h-LC_{50} = 7.40$	Gu et al.,
28 ± 0.5 °C	mg/L	2020
14 h light per day	9/ =	
7.08 - 72.43 - 700.43 µg/L (mean	Free-swimming total distance:	(Klimisch 2)
measured conc.)	tended to decline in a dose-	
TUNEL staining	dependent manner (sign at 70 and	
	700 µg/L) 300 µg/L CPF (positive control):	
	same effect	
	Oxidative stress in 3 dpf and 6 dpf	
	larvae: decrease of activities of	
	SOD, increase of activity of MDA	
	(dose-dependent)	
	Apoptosis in larvae brain: increased number of death cell in larvae brain	
	with increasing BPF conc. (sign at	
	70 and 700 µg/L)	
	Decreased expressions of genes	
	regulating neurodevelopment (dose-	
	dependent)	
	Histopathological abnormalities in brain (sign at 700 μg/L)	
Danio rerio Embryo (< 4 hpf)	Survival:	Lee et al.,
Until 120 hpf	NOEC = 2.0 (BPA) or 10 (BPF) mg/L	2019
DMSO < 0.1%	The state of the balls	(1/11:1:1:-2)
Carbon-filtered tap water 25 ± 1 °C	Time-to-hatch: LOEC = 2.0 (BPA) or 0.08 (BPF)	(Klimisch 2)
14 h light/day	mg/L	
96-well plates	1119/ =	
Observation of survival, hatching,	Hatchability:	
malformation: 1 larva per well, 6 wells	No sign. effect for BPF; NOEC = 2.0	
per replicate, 4 replicates per	mg/L for BPA	
treatment at 0.4, 2, and 10 mg/L for both BPA and BPF;	Sign. increase of T3:	
Measurement of THs in larval fish: 160	At 0.4 mg/L BPA	
larvae per replicate, three replicates	Sign increase of T4:	
per treatment in a glass beaker (200	At 2.0 mg/L for BPF	
mL volume) at 0.08, 0.4, and 2 mg/L	Transportational shares as to TU	
for both BPA and BPF at 0.4 and 2.0 mg/L for both BPA and BPF;	Transcriptional changes in TH regulating genes:	
Transcriptional analysis: mass chamber	Up-regulated hematopoietically	
(50 mL of media/batch) with 25 per	expressed homeobox (hhex),	
replicates and three replicates per	transthyretin (ttr) and ridine	
treatment at 0.08, 0.4, and 2 mg/L for	diphosphate glucuronosyltransferase	
both BPA and BPF	1 ab (ugt1ab) genes by BPA, BPF, and BPS.	
Danio rerio (wild-type AB strain and	Hatching success (measured at 55	Coumailleau
transgenic lines Tg(cyp19a1b:GFP))	hpf):	et al., 2020
eggs collected 2hpf	BPÁ sign. at both conc.	
28.5 °C	BPF no sign. effect	(Klimisch 2)

80 embryos per treatment Glass flask (100 mL medium)	BPAF sign. at 1 µM BPS sign. at both conc.	
28.5°C ± 0.5°C	Br 3 sign. at both conc.	
14 hour light per day	Locomotor activity (using	
DMSO, EE2 (1nM = positive control) BPA, BPS, BPF, BPAF, BPAP: 1 and 0.1	ZEBRALAB): No sign effect on swimming	
μM	distance, but reduction of swimming	
	speed over time for all substances	
	tested. BPAF showed as ignificantly	
	reduced locomotor activity following 6 days of exposure.	
Danio rerio (AB strain)	Sign increased 1L-1ß and TNF-a	Wang et al.,
40 fish (20 males + 20 females) for	levels (indicating intestinal	2021
one group	inflammation) ELISA analysis showed induced	
2 replicates per conc. 14 days exposure	oxidative damage and inflammatory	(Klimisch 2)
1, 10, 100, 1000 μg/L	response	(
≤ 0.01% DMSO	Different changes in microbial	
Semi-static (24h renewal) ELISA analysis	community of zebrafish intestine	
RNA-seq analysis		
qRT-PCR analysis		
DNA extraction (In vitro and) in vivo zebrafish	BPA, BPF, BPS: induced GFP in a	Le Fol et al.,
(21. VIGO GIIG) III VIVO ZEDIGIISII	concentration-dependent manner.	2017
quantifying the expression of brain	BPS only partially induced brain	
aromatase using a transgenic <i>cyp19a1b</i> -GFP zebrafish	aromatase at the highest tested concentrations (>30 µM) while BPA	(Klimisch 2)
embryo assay	and BPF strongly induced GFP, in an	
	ER-dependent manner, at 1–10 μM.	
	Additionally, BPF strongly induced	
	vitellogenin synthesis in adult male zebrafish.	
Danio rerio	BPF, BPS, not BPA: reduced	Wu and
21 d exposure to BPA, BPF, or BPS	swimming performance – no	Seebacher,
30 μg/L – 0.005% v/Vv ethanol Acclimatisation or acute test (18°C and	interactions between bisphenol exposure and acclimatisation	2021
28 °C)	exposure and decimalisation	(Klimisch 2)
14 h light per day		
9-15 fish per treatment Acute toxicity testing to algae,	Daphnia 48h-EC50= 7.3 mg BPA/L	Tišler et al.,
daphnia, fish + chronic	or 8.7 mg BPF/L	2016
	Zebrafish 72h-EC50 (hatching inh)=	
	4.0 mg BPA/L or 6.8 mg BPF/L Zebrafish 48h LC50= 15.9 mg	(Klimisch 2)
	BPA/L or BPF not determined as not	
	as toxic	
Zebrafish embryos and larvae	reduced amount of eggs and hatching rate	Ren et al., 2017
		201/
		(Klimisch 2)
Danio rerio	$96h-LC_{50} = 1.6 \cdot 10^6 \mu M$	Han et al., 2021
96h- acute toxicity + gene expression 1, 2, 4, 6, 8, 10, 12, 14 mg·L ⁻¹		2021
		(Klimisch 2)
Zebrafish (wild type AB)	60d: no sign difference in fish body	Qiu et al.,
Juvenile 10 – 15 g	weight, survival rates, body length compared to control	2019
5 L glass tanks	compared to control	(Klimisch 2)
5 fish per tank, 4 replicates	Increased expression of	
0.005% DMSO, blank, 0.1, 1, 10, 100, and 1000 μg/L	reproductive neuroendocrine-related genes and increased levels of	
	l actica and increased levels of	
Analytical monitoring	hormones in the zebrafish brain as	

	T 11	
Semi-static	well as increased levels of	
100 µg/L BPA or BPF for 60 d -	vitellogenin in liver	
survival rate, body weight, body		
length, HSI		
Cyprinus carpio	60d: no sign difference in fish body	Qiu et al.,
Juvenile	weight, survival rates, body length	2018
10 - 15 g	compared to control	2010
5 L glass tanks	30pa. 34 to 30	(Klimisch 2)
5 fish per tank, 4 replicates	HSI sign increased at 1000 µg/L BPF	(**************************************
0.005% DMSO, blank, 0.1, 1, 10, 100,	1 3,	
and 1000 µg/L	Oxidative stress at 100 and 1000	
Analytical monitoring	μg/L BPF	
Semi-static		
	BPF similar effects as BPA on	
100 μg/L BPA or BPF for 60 d –	immune modulation	
survival rate, body weight, body		
length, HSI,		
Zebrafish (wild-type AB) embryos (2	Pigmentation reduction (sign less	Mu et al.,
hpf)	melanin in eyes, yolk sac,	2019
Developmental tox test:	notochord)	///!:==!==! 2\
24-well plates	Cine and number of real-reading	(Klimisch 2)
Blank -solvent control (Acetone) -	Size and number of melanocytes	
0.005 - 0.5 - 5 mg/L; 3 replicates; 96h expo	sign reduced in embryos of 5 mg/L BPF treatment	
Transcriptomic test: 1-L beakers; 3	DPF treatment	
replicates; Blank -solvent control	Developmental defects: abnormal	
(Acetone) – 0.005 – 0.5 mg/L; 50	spontaneous movements at 0.5 and	
embryos per beaker; at 48 hpf: 25	5.0 mg/L; decreased heart rate	
hatched larvae per replicate evaluated	hatch inhibition sign at 5.0 mg/L	
qPCR validation: 0.0005, 0.5 and	materi ministrioni signi at sio mig, z	
5.0 mg/L in 1L-beakers; 100 embryos;	Sign. effect on motor neuron	
3 replicates; at 48 hpf: 25 hatched	development at 0.5 or 5.0 mg/L	
larvae per replicate evaluated	,	
analytical verification of conc.		
(measured less than 20% different		
from nominal)		
Zebrafish (<i>Danio rerio</i> , AB strain)	No solvent effect	Mu et al.,
Acute tox test: embryos 1.55-1.7 hpf		2018
in 24-well plates with 2 mL exposure	$96h-LC_{50} = 19.59 \text{ mg/L}$	
solution in triplicate		
Expo: 4 days	Sign inhibition of embryo hatching	
Semi-static (renewal every 24h)	ratio: 15% at 5 mg/L and 65% at	
14 hours light per day 26°C	10 mg/L at 72 hpf	
Developmental Toxicity Test:	Pigmentation of eyes, yolk sac,	
embryos 1.55-1.7 hpf in 24-well plates	notochord of embryos sign reduced	
with 2 mL exposure solution in	(35% at 1 mg/L; 57% at 5 mg/L;	
triplicate (control, 0.1, 0.5, 1.0 mg/L)	82% at 10 mg/L)	
Estrogenic Activity Test: about 100	= dc 10g/ =/	
embryos in 1-L-beakers with 500 mL		
exposure solution in triplicate (0.2, 2,		
10 mg/L)		
Expo: 2 hpf to 96 hpf		
60 embryos per replicate collected		
ELISA tests: 30 embryos per sample		
homogenized with saline on ice		
Gene Expression Analysis: Total RNA		
extracted from 30 embryos per sample		
+ qPCR validation		
Chemical confirmation of substances:		
UPLC-MS		
Danio rerio (Zebrafish, AB strain)	Increase of T3 and decrease of T4	Huang et al.,

larvae	contents, increased ratios of T3/T4	2016
0.01% DMSO, 0.2, 2, 20, and 200 μg/L	TSH content sign induced in	
2 hpf to 144 hpf	concentration-dependent manner	(Klimisch 2)
500 embryos; 6 replicates	Increased gene transcription of	
LC-MS/MS	dio2, crh and nis and tg	

Table 7: Available in vivo data with amphibia on BPF

Method	Results	Reference
Xenopus laevis	In the presence of T3, higher	Zhu et al.,
T3-induced metamorphosis assay	concentrations of BPF (100–10000	2018
Tadpoles stage 52 in glass tanks (9 per	nM) antagonized T3-induced TH-	
tank); 3 replicates	response gene transcription and	(Klimisch 2)
0.001% DMSO	morphological changes in a	
10, 100, 1000, 10000 nM in absence or	concentration-dependent manner,	
presence of 1 nM T3 22 ± 1 °C	whereas 10 nM BPF exerted stimulatory effects on T3-induced	
22 ± 1 °C 24h expo	integral metamorphosis, displaying	
In vitro tail assay	TH signaling disrupting effects with	
Spontaneous metamorphosis assay	complicated concentration-response	
Histological examination	relationships	
This cological examination	T clade of lot lips	
	In the absence of T3, BPF inhibited	
	development at metamorphic	
	climax, but promoted pre- and pro-	
	metamorphic development,	
	displaying a developmental stage-	
	dependent manner	
	aganistic actions of DDF on Noteh	
	agonistic actions of BPF on Notch signaling in the intestine	
Pelophylax nigromaculatus tadpoles	BPA, BPF, BPS induced TH-response	Zhang et al.,
(Gosner stage 27)	gene transcription in the tadpoles	2018
TH-response gene transcription assay	In presence of T3 altered T3-	2010
0.01–10 µM BPA, BPS, or BPF in the	induced gene transcription in	(Klimisch 2)
presence or absence of 0.2 nM T3 for	biphasic concentration-response	` ,
48 h	manner	
Three replicate test beakers		
Xenopus laevis	TH-response gene expression: BPA	Niu et al.,
NF 52 stage tadpoles	increases like st3, dio3 and thibz	2021
Semi-static test system	expression in brain – BPF	(1/limpig = 1- 2)
T3-induced (1nM) Xenopus	upregulated <i>dio3</i> but not the other; co-exposure: BPA antagonised T2-	(Klimisch 2)
metamorphosis assay 10, 100, and 1000 nM BPA or BPF in	induced upregulation of tgm2, thibz,	
absence or presence of 1 nM T3	st3, and dio3 expression - In the	
Glass tanks (4L water)	presence of T3, BPF exhibited more	
9 tadpoles per tank with three	remarkable biphasic effects on T3-	
replicates	induced expression of all the test	
0.001% 8v/v) DMSO	TH-response genes except for <i>tgm2</i> .	
96h test duration	Brain morphology:	
	BPA and BPF antagonised T3-	
	induced brain remodelling	

3.3.2.1. Fish data

There is a clear indication of an estrogenic endocrine MoA in the zebrafish studies. One study observed a significant dose-dependent increase in VTG levels in male zebrafish after 7 days of exposure to 0.1 μ M BPF in the water (Le Fol et al., 2017). Another study also observed a significant increase in VTG in both male and female zebrafish, after 60 days exposure to 0.0001 mg/L water (Yang et al., 2018). In addition, the endocrine MoA is

supported by an observation of increased estradiol levels in both juvenile and adult male and female zebrafish, at doses of 0.01 mg/L water (juvenile) and 0.1 mg/L water (adult) (Yang et al., 2017 and 2018). Furthermore, Qui et al. (2019) showed that long-term exposure to low and environmentally relevant levels of BPF in vivo leads to increased expression of reproductive neuroendocrine-related genes and increased levels of hormones in the zebrafish brain as well as increased levels of VTG in liver. The authors suggest that chronic BPF exposure affects regulation of the reproductive neuroendocrine system through activation of the estrogen receptor.

3.3.2.2. Amphibian data

In vivo mechanistic data from *Xenopus laevis* and *Pelophylax nigromaculatus* tadpoles show that BPF exerts some activity on the HPT axis. Zhou et al. (2018) found that in the absence of T3, BPF inhibited development of the tadpoles at metamorphic climax, but promoted pre- and pro-metamorphic development, displaying a developmental stage-dependence. The study performed by Zhang et al. (2018) on *Pelophylax nigromaculatus* tadpoles showed that BPA, BPF, BPS induced TH-response gene transcription in the tadpoles. In presence of T3 altered T3-induced gene transcription was observed in a biphasic concentration-response manner. Niu et al. (2021) found effects of BPF on thyroid specific gene expression as well as antagonised T3-induced brain remodelling in *X. laevis*.

3.3.2.3. Conclusion

The available fish data clearly and consistently show an estrogenic and/or anti-androgenic mode of action of BPF. The effect concentrations are in the low mg/L range (0.01-1 mg/L) and hence comparable to those observed for BPA and 4-nonylphenol⁴ which are already identified ED substances based on their estrogenic activity.

This observed *in vivo* endocrine activity fits to the observed endocrine effects of BPF in the available *in vitro* data.

Furthermore, BPF shows activity on the HPT axis, which also fits to the effects observed in the available *in vitro* data and effects reported for BPA.

3.3.3. In vivo adverse effect data

3.3.3.1. Fish data

Adverse and population relevant effects in fish were observed in the study of Yang et al. (2017). In this study *Danio rerio* (AB strain, 6 males and 6 females in two replicates) was exposed in a semi static set up to 0.001, 0.01, 0.1, and 1 mg/L BPF for 21 days following OECD guidelines 229 and 230. Histological analysis was performed via microscopy, hormone measurements (testosterone and E2) via ELISA were included and a gene expression analysis (vtg1) was done. Egg production in F0, hatching rate and survival rate of F1 were significantly reduced at 1 mg/L of BPF. Hence, clear adverse effects on reproduction could be shown.

HSI at 1 mg/L was significantly higher in males and GSI was found to be significantly decreased in females and males at 1 mg/L. The histological examination revealed a decrease in the number of early sperm stages and an enlargement of interstitial space in

⁴ Annex XV dossier containing the proposal for identification of 4-nonylphenol, branched and linear [...] as a Substance of Very High Concern based on its environmental ED properties: https://echa.europa.eu/de/registry-of-svhc-intentions/-/dislist/details/0b0236e180e4ba35

male fish. In female fish a lower proportion of later follicular stages was observed and no intersex individuals could be detected. Significant differences in concentrations of steroid hormones were observed with a decrease in testosterone concentration in the full body homogenate of the 0.001, 0.1 and 1 mg/L treatment. A significant increase of E2 was observed at concentrations of 0.1 and 1 mg/L BPF. The gene expression analysis performed in liver homogenate revealed a significant increased vtg1 expression in males (dose-dependent), but not in females.

Furthermore, the long-term fish study performed by Yang et al. (2018) on zebrafish showed adverse and population relevant effects. In this study, 500 eggs were exposed in triplicate to 0, 1, 10, 100, or 1000 μ g/L of BPF under semi-static conditions. At 10 dpf the larvae were transferred to 10 L glass aquaria, and exposure to BPF was continued to 60 dpf. Histology of the gonads was performed at the end of the test, and E2 and testosterone concentrations in full body homogenate were measured via ELISA. Real-time PCR was performed for gene expression analysis (vtg1 in liver). Adverse effects were seen at 100 and 1000 μ g/L BPF since histology of the gonads revealed that there is a significant increase in intersex fish of 10 and 22%, respectively. Simultaneously, all males developed ovo-testes and showed an abnormal development. Additionally, there was a trend towards a female-biased sex ratio with a statistically significant shift of the sex ratio towards females at 100 μ g/L.

Hormone measurements showed a decrease in testosterone concentration significant from 10 μ g/L onwards and an increase of E2 also significant from \geq 10 μ g/L. Liver VTG expression increased significantly from 20 dpf in the \geq 10 μ g/L exposure groups.

3.3.3.2. Conclusion

Two fish studies performed with *Danio rerio* clearly show consistent adverse effects on the reproductive capacity of the animals after exposure to BPF. The study performed by Yang et al. followed an OECD TG 229 (fish short-term reproduction assay) and demonstrated adverse effects on egg production in F0 females as well as on hatching and survival of the larvae of the F1 generation. The same study shows effects for an estrogenic and/or anti-androgenic mode of action in absence of indications for unspecific systemic toxicity. Hence, according to OECD GD 150 this study can be used to conclude on ED mediated adversity. The long-term fish study of Yang et al. (2018) investigated the sexual development of zebrafish. Here, clear adverse effects could be shown on gonadal development and on the sex ratio. The further endpoints investigated in this study clearly point to an underlying estrogenic and/or anti-androgenic mode of action of these adverse effects. Indications for systemic toxicity that could have triggered the observed effects on sexual development were not seen in this study. According to OECD GD 150, the effects observed in this study are also conclusive for ED mediated adverse and population relevant effects.

3.4. Conclusion regarding ED properties relevant for environment

3.4.1. Adverse effects relevant for ED identification

As described above, the available data for BPF show clear and consistent adverse and population relevant effects on reproduction and sexual development in two zebrafish studies. The observed effects fit to an estrogenic and/or anti-androgenic mode of action and no indications of further non-ED mediated pathways were found in the two studies.

3.4.2. Endocrine activity

The available *in vitro* as well as *in vivo* mechanistic data clearly and consistently demonstrate an estrogenic and anti-androgenic activity of BPF.

Additionally, in vitro and in vivo data from fish and amphibians point to an interference of

BPF with the HPT axis.

Two *in vitro* studies also show an interference with testosterone production in cellular assays.

3.4.3. Plausible link between adverse effects and endocrine activity

BPF may have multiple modes of endocrine action (estrogenic, anti-androgenic, thyroidal activity and interference with steroidogenesis) that might interact and are difficult to distinguish from each other. However, the estrogenic and/or anti-androgenic effects of BPF in fish are consistently observed in the available studies showing significant effects on egg production, hatching and survival of F1 larvae, sex ratio and gonadal development. Estrogenic and anti-androgenic modes of action are well known to be involved in the regulation of sexual development and reproduction (AOP Wiki including examples therein⁵, e.g. AOP 345). Adverse effect such as feminization of fish, fertilization success, ability to produce viable offspring and gonadal development and their link to an estrogenic and/or anti-androgenic mode of action have been reviewed i.a. by Jobling et al. (2002 and 1998), Miller et al. (2012).

Considering the observed concomitant decrease in plasma testosterone levels and the increase in plasma estradiol levels as well as the increase in VTG levels and gene expression in male fish, the link between these endocrine activities and the observed adverse effects on fish is highly plausible.

3.4.3.1. Other supporting evidence

In vitro and *in vivo* data from fish and amphibians point to an interference of BPF with the hypothalamic-pituitary-thyroid (HPT) axis.

Additionally, the available human health data support the conclusion for an estrogenic and/or antiandrogenic activity of BPF, even though the reliability of these studies have not been assessed here. Studies on rats show consistently a decrease in serum testosterone levels and a decrease in sperm motility in the offspring of treated female rats. An increase in uterus weight was observed in juvenile rats.

The link between the observed effects and the specific estrogenic and/or anti-androgenic activity of BPF is further supported by the analogy of BPF to BPA and BPA. The data available for BPA and BPB, both of which share very similar chemicals structures compared to BPF, show well defined adverse effects and modes of action that fit to an estrogenic mode of action in fish. Based on these data BPA and BPB have been identified already as SVHC due to its endocrine disrupting properties in the environment.

3.4.4. Conclusion on ED properties

Overall, BPF has estrogen agonistic properties and induces adverse effects in fish that are plausibly mediated by this endocrine activity.

Furthermore, *in vivo* and *in vitro* evidence shows that BPF has androgen antagonistic properties. This endocrine activity could also plausibly contribute to the observed adverse

_

⁵ https://aopwiki.org/aops

effects on reproduction and sexual development in fish.

The effects observed in fish are relevant for the environment as an effect on the reproductive function and the sexual development can have consequences at a population level. Therefore, there is scientific evidence to conclude that BPF fulfils the definition of an endocrine disruptor in the environment.

References

Aykut, H., & Kaptaner, B. (2021). In vitro effects of bisphenol F on antioxidant system indicators in the isolated hepatocytes of rainbow trout (Oncorhyncus mykiss). Molecular biology reports.

Catron TR, Keely SP, Brinkman NE, Zurlinden TJ, Wood CE, Wright JR, Phelps D, Wheaton E, Kvasnicka A, Gaballah S, Lamendella R, Tal T. Host Developmental Toxicity of BPA and BPA Alternatives Is Inversely Related to Microbiota Disruption in Zebrafish. Toxicol Sci. 2019 Feb 1;167(2):468-483. doi: 10.1093/toxsci/kfy261. PMID: 30321396.

Coumailleau P, Trempont S, Pellegrini E, Charlier TD. Impacts of bisphenol A analogues on zebrafish post-embryonic brain. J Neuroendocrinol. 2020 Aug;32(8):e12879. doi: 10.1111/jne.12879. Epub 2020 Aug 4. PMID: 32749037.

ECHA & EFSA, with the technical support of the Joint Research Centre (JRC), Niklas Andersson, Maria Arena, Domenica Auteri, Stefania Barmaz, Elise Grignard, Aude Kienzler, Peter Lepper, Alfonso Maria Lostia, Sharon Munn, Juan Manuel Parra Morte, Francesca Pellizzato, Jose Tarazona, Andrea Terron, and Sander Van der Linden. 2018. "Guidance for the identification of endocrine disruptors in the context of Regulations (EU) No 528/2012 and (EC) No 1107/2009." EFSA Journal 16 (6):e05311. doi: doi:10.2903/j.efsa.2018.5311.

Eladak S, Grisin T, Moison D, Guerquin MJ, N'Tumba-Byn T, Pozzi-Gaudin S, Benachi A, Livera G, Rouiller-Fabre V, Habert R. A new chapter in the bisphenol A story: bisphenol S and bisphenol F are not safe alternatives to this compound. Fertil Steril. 2015 Jan;103(1):11-21. doi: 10.1016/j.fertnstert.2014.11.005. Epub 2014 Dec 2. PMID: 25475787.

Gu J, Wu J, Xu S, Zhang L, Fan D, Shi L, Wang J, Ji G. Bisphenol F exposure impairs neurodevelopment in zebrafish larvae (Danio rerio). Ecotoxicol Environ Saf. 2020 Jan 30;188:109870. doi: 10.1016/j.ecoenv.2019.109870. Epub 2019 Nov 1. PMID: 31683046.

Han, Y., Fei, Y., Wang, M., Xue, Y., Chen, H., & Liu, Y. (2021). Study on the Joint Toxicity of BPZ, BPS, BPC and BPF to Zebrafish. Molecules, 26.

Huang GM, Tian XF, Fang XD, Ji FJ. Waterborne exposure to bisphenol F causes thyroid endocrine disruption in zebrafish larvae. Chemosphere. 2016 Mar;147:188-94. doi: 10.1016/j.chemosphere.2015.12.080. Epub 2016 Jan 7. PMID: 26766355.

Jobling, Susan; Nolan, Monique; Tyler, Charles R.; Brighty, Geoff; Sumpter, John P. Environmental Science and Technology (1998), 32 (17), 2498-2506CODEN: ESTHAG; ISSN:0013-936X (American Chemical Society).

Jobling, S.; Coey, S.; Whitmore, J. G.; Kime, D. E.; Van Look, K. J. W.; McAllister, B. G.; Beresford, N.; Henshaw, A. C.; Brighty, G.; Tyler, C. R.; Sumpter, J. P. Biology of Reproduction (2002), 67 (2), 515-524CODEN: BIREBV; ISSN:0006-3363. (Society for the Study of Reproduction).

JRC, Munn S, Goumenou M. Key scientific issues relevant to the identification and characterisation of endocrine disrupting substances - Report of the Endocrine Disrupters Expert Advisory Group. EUR 25919. Luxembourg (Luxembourg): Publications Office of the European Union; 2013. JRC79981.

Kitamura S, T. Suzuki, S. Sanoh, R. Kohta, N. Jinno, K. Sugihara, S. Yoshihara, N. Fujimoto, H. Watanabe, S. Ohta. Comparative study of the endocrine-disrupting activity of bisphenol A and 19 related compounds. Toxicol. Sci., 84 (2005), pp. 249-259

Klimisch, H J; Andreae, M; Tillmann, U, A Systematic Approach for Evaluating the Quality of Experimental Toxicological and Ecotoxicological Data, Regulatory Toxicology and Pharmacology 1997, 25 1: 1–5. doi:10.1006/rtph.1996.1076

Lee S, Kim C, Shin H, Kho Y, Choi K. Comparison of thyroid hormone disruption potentials by bisphenols A, S, F, and Z in embryo-larval zebrafish. Chemosphere. 2019 Apr;221:115-123. doi: 10.1016/j.chemosphere.2019.01.019. Epub 2019 Jan 4. PMID: 30639807.

Le Fol V, S. Aït-Aïssa, M. Sonavane, J. Porcher, P. Balaguer, J. Cravedi, D. Zalko, F. Brion. In vitro and in vivo estrogenic activity of BPA, BPF and BPS in zebrafish-specific assays. Ecotoxicol. Environ. Saf., 142 (2017), pp. 150-156

Liu J, Zhang L, Lu G, Jiang R, Yan Z, Li Y. Occurrence, toxicity and ecological risk of Bisphenol A analogues in aquatic environment - A review. Ecotoxicol Environ Saf. 2021 Jan 15;208:111481. doi: 10.1016/j.ecoenv.2020.111481. Epub 2020 Oct 23. PMID: 33120264.

Miller, HD, Clark BW, Hinton DE, Whitehead A, Martin S, Kwok KW, Kullman SW. Anchoring Ethinylestradiol Induced Gene Expression Changes with Testicular Morphology and Reproductive Function in the Medaka. PLoS ONE 2012, 7(12): e52479. DOI: 10.1371/journal.pone.0052479.

- Moreman J, O. Lee, M. Trznadel, A. David, T. Kudoh, C.R. Tyler. Acute toxicity, teratogenic, and estrogenic effects of Bisphenol A and its alternative replacements Bisphenol S, Bisphenol F, and Bisphenol AF in zebrafish embryo-larvae. Environ. Sci. Technol., 51 (2017), pp. 12796-12805
- Mu, X., Huang, Y., Li, X., Lei, Y., Teng, M., Li, X., Wang, C., Li, Y., 2018. Developmental effects and estrogenicity of bisphenol A alternatives in a zebrafish embryo model. Environ. Sci. Technol. 52, 3222–3231. https://doi.org/10.1021/acs.est.7b06255.
- Mu X, Liu J, Yuan L, Yang K, Huang Y, Wang C, Yang W, Shen G, Li Y. The mechanisms underlying the developmental effects of bisphenol F on zebrafish. Sci Total Environ. 2019 Oct 15;687:877-884. doi: 10.1016/j.scitotenv.2019.05.489. Epub 2019 Jun 10. PMID: 31412491.
- Niu Y, Zhu M, Dong M, Li J, Li Y, Xiong Y, Liu P, Qin Z. Bisphenols disrupt thyroid hormone (TH) signaling in the brain and affect TH-dependent brain development in Xenopus laevis. Aquat Toxicol. 2021 Aug;237:105902. doi: 10.1016/j.aquatox.2021.105902. Epub 2021 Jun 24. PMID: 34218114.
- OCDE (2018), Revised Guidance Document 150 on Standardised Test Guidelines for Evaluating Chemicals for Endocrine Disruption, OECD Series on Testing and Assessment, n° 150, Éditions OCDE, Paris, https://doi.org/10.1787/9789264304741-en .
- Park C, Song H, Choi J, Sim S, Kojima H, Park J, Iida M, Lee Y. The mixture effects of bisphenol derivatives on estrogen receptor and androgen receptor. Environ Pollut. 2020 May;260:114036. doi: 10.1016/j.envpol.2020.114036. Epub 2020 Jan 21. PMID: 31995776.
- Pelch KE, Li Y, Perera L, Thayer KA, Korach KS. Characterization of Estrogenic and Androgenic Activities for Bisphenol A-like Chemicals (BPs): In Vitro Estrogen and Androgen Receptors Transcriptional Activation, Gene Regulation, and Binding Profiles. Toxicol Sci. 2019 Aug 6;172(1):23–37. doi: 10.1093/toxsci/kfz173. Epub ahead of print. PMID: 31388671; PMCID: PMC6813750.
- Punt A, Aartse A, Bovee TFH, Gerssen A, van Leeuwen SPJ, Hoogenboom RLAP, Peijnenburg AACM. Quantitative in vitro-to-in vivo extrapolation (QIVIVE) of estrogenic and anti-androgenic potencies of BPA and BADGE analogues. Arch Toxicol. 2019 Jul;93(7):1941-1953. doi: 10.1007/s00204-019-02479-6. Epub 2019 May 20. PMID: 31111190.
- Qui, Wenhui; Fang, Meijuan; Liu, Jingyu; Fu, Caixia; Zheng, Chunmiao; Chen, Bei and Wang, Ke-Jian (2019): In vivo actions of Bisphenol F on the reproductive neuroendocrine system after long-term exposure in zebrafish. Science of The Total Environment. Volume 665, 15 May 2019, Pages 995-1002. DOI: 10.1016/j.scitotenv.2019.02.154.
- Qiu W, Zhan H, Tian Y, Zhang T, He X, Luo S, Xu H, Zheng C. The in vivo action of chronic bisphenol F showing potential immune disturbance in juvenile common carp (Cyprinus carpio). Chemosphere. 2018 Aug;205:506-513. doi: 10.1016/j.chemosphere.2018.04.105. Epub 2018 Apr 23. PMID: 29705641.
- Ren W.J., Z. Wang, L. Wang, X.H. Yang, J.N. Liu, et al. Effects of Bisphenol A and its analogues on zebrafish embryos and larvae Asian J. Ecotoxicol., 12 (2017), pp. 184-192 (in Chinese)
- Šauer P, Švecová H, Grabicová K, Gönül Aydın F, Mackul'ak T, Kodeš V, Blytt LD, Henninge LB, Grabic R, Kocour Kroupová H. Bisphenols emerging in Norwegian and Czech aquatic environments show transthyretin binding potency and other less-studied endocrine-disrupting activities. Sci Total Environ. 2021 Jan 10;751:141801. doi: 10.1016/j.scitotenv.2020.141801. Epub 2020 Aug 19. PMID: 32861950.
- Tišler T, A. Krel, U. Gerželj, B. Erjavec, M.S. Dolenc, A. Pintar. Hazard identification and risk characterization of Bisphenols A, F and AF to aquatic organisms . Environ. Pollut., 212 (2016), pp. 472-479
- Ullah A, Pirzada M, Jahan S, Ullah H, Shaheen G, Rehman H, Siddiqui MF, Butt MA. Bisphenol A and its analogs bisphenol B, bisphenol F, and bisphenol S: Comparative in vitro and in vivo studies on the sperms and testicular tissues of rats. Chemosphere. 2018 Oct;209:508-516. doi: 10.1016/j.chemosphere.2018.06.089. Epub 2018 Jun 19. PMID: 29940534.
- van Leeuwen SP, Bovee TF, Awchi M, Klijnstra MD, Hamers AR, Hoogenboom RL, Portier L, Gerssen A (2019) BPA, BADGE and analogues: a new multi-analyte LC-ESI-MS/MS method for their determination and their in vitro (anti)estrogenic and (anti)androgenic properties. Chemosphere 221:246–253. https://doi.org/10.1016/J.CHEMOSPHERE.2018.12.189
- Wang Y, Wang B, Wang Q, Liu Y, Liu X, Wu B, Lu G. Intestinal toxicity and microbial community disorder induced by bisphenol F and bisphenol S in zebrafish. Chemosphere. 2021 Oct;280:130711. doi: 10.1016/j.chemosphere.2021.130711. Epub 2021 Apr 30. PMID: 34162083.
 - Wu NC, Seebacher F. Bisphenols alter thermal responses and performance in zebrafish (Danio

rerio). Conserv Physiol. 2021 Jan 16;9(1):coaa138. doi: 10.1093/conphys/coaa138. PMID: 33505703; PMCID: PMC7816798.

Yang Q, Yang X, Liu J, Chen Y, Shen S. Effects of exposure to BPF on development and sexual differentiation during early life stages of zebrafish (Danio rerio). Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology. 2018 August; 210: 44-46. https://doi.org/10.1016/j.cbpc.2018.05.004

Yang Q, Yang X, Liu J, Ren W, Chen Y, Shen S. Effects of BPF on steroid hormone homeostasis and gene expression in the hypothalamic-pituitary-gonadal axis of zebrafish. Environ Sci Pollut Res Int. 2017 Sep;24(26):21311-21322. doi: 10.1007/s11356-017-9773-z. Epub 2017 Jul 25. PMID: 28741210.

Yuan L, Qian L, Qian Y, Liu J, Yang K, Huang Y, Wang C, Li Y, Mu X. Bisphenol F-Induced Neurotoxicity toward Zebrafish Embryos. Environ Sci Technol. 2019 Dec 17;53(24):14638-14648. doi: 10.1021/acs.est.9b04097. Epub 2019 Nov 25. PMID: 31702913.

Zhang YF, Ren XM, Li YY, Yao XF, Li CH, Qin ZF, Guo LH. Bisphenol A alternatives bisphenol S and bisphenol F interfere with thyroid hormone signaling pathway in vitro and in vivo. Environ Pollut. 2018 Jun;237:1072-1079. doi: 10.1016/j.envpol.2017.11.027. Epub 2017 Nov 13. PMID: 29146198.

Zhu M, Chen XY, Li YY, Yin NY, Faiola F, Qin ZF, Wei WJ. Bisphenol F Disrupts Thyroid Hormone Signaling and Postembryonic Development in Xenopus laevis. Environ Sci Technol. 2018 Feb 6;52(3):1602-1611. doi: 10.1021/acs.est.7b06270. Epub 2018 Jan 25. PMID: 29323886.